



Review

AEROGENIC POLLUTANTS AS RISK FACTORS CAUSING DEVELOPMENT OF CARDIO-METABOLIC PATHOLOGY (REVIEW)

A.E. Nosov¹, A.S. Baydina¹, O.Yu. Ustinova^{1,2}¹Federal Scientific Center for Medical and Preventive Health Risk Management Technologies, 82 Monastyrskaya Str., Perm, 614045, Russian Federation²Perm State National Research University, 15 Bukireva Str., Perm, 614990, Russian Federation

Ambient air pollution causes approximately 3.3 million untimely deaths annually (2.1 deaths due to ischemic heart disease and 1.1 million deaths due to stroke). Mortality caused by ambient air pollution is higher than mortality due to such traditional risk factors as smoking, obesity, and elevated dextrose contents in blood. Relative risk of mortality amounts to 1.26 (95 % CI 1.08–1.47) in cities with the highest air pollution against those where air pollution is the lowest. Occupational exposure to various chemical air pollutants can cause more than 1 million untimely deaths all over the world but its contribution to prevalence of cardiovascular diseases has not been determined sufficiently. Aero-genic pollutants are quite variable in their chemical structure and include both particulate matter (PM for short) and gaseous matter. The American Heart Association and the European Society of Cardiology consider PM_{2.5} to be a risk factor causing cardiovascular diseases. This analytical review presents data on effects produced by aero-genic pollutants on development of cardio-metabolic pathology and population mortality due to vascular and metabolic diseases (arterial hypertension, atherosclerosis and ischemic heart disease, heart rhythm disturbances, and type 2 diabetes mellitus). There are also data on mechanisms of pathogenetic influence exerted by aero-genic pollutants on development of such diseases including generation of anti-inflammatory and oxidative mediators and their release into blood flow; developing imbalance in the autonomic nervous system with prevailing activity of the sympathetic nervous system and disrupted heart rate variability; direct introduction of aero-genic pollutants from the lungs into blood flow with developing direct toxic effects. We have also analyzed literature data on protective effects produced by reduction in ambient air pollution on prevalence of cardiovascular pathology.

Key words: aero-genic pollutants, airborne particulate matter, persistent organic pollutants, cardiovascular pathology.

Cardiovascular pathology remains a leading cause of mortality among employable population all over the world since 17 million untimely deaths result from it annually; 3.3 million out of them are associated with ambient air pollution with technogenic chemicals [1, 2]. Effects produced by ambient air pollution occupy the 6th rank place among mortality factors being ahead of such traditional risk factors as smoking (2.48 million), obesity (2.85 million), and elevated glucose contents in blood (2.84 million) [3]. Only arterial hypertension (AH) is more significant as a

risk factor of cardiovascular mortality than ambient air pollution [2]. According to E. Braunwald, relative risk of mortality amounts to 1.26 (95 % CI 1.08–1.47) in the most polluted cities against the least polluted ones [4]. Occupational exposure to various chemical air pollutants can cause more than 1 million untimely deaths all over the world but its contribution to cardiovascular mortality hasn't been well-determined yet [5].

Ambient air pollutants are quite heterogeneous in their chemical structure and include both particulate matter (PM for short) and

© Nosov A.E., Baydina A.S., Ustinova O.Yu., 2021

Alexander E. Nosov – Candidate of Medical Sciences, Head of In-patient Clinic (Therapeutic Work-related Pathology Department) (e-mail: nosov@fcrisk.ru; tel.: +7 (342) 219-87-36; ORCID: <http://orcid.org/0000-0003-0539-569X>).

Anastasia S. Baydina – Candidate of Medical Sciences, cardiologist at the Consulting and Polyclinic Department (e-mail: anastasia_baidina@mail.ru.ru; tel.: +7 (342) 219-87-36; ORCID: <http://orcid.org/0000-0003-3131-5868>).

Olga Yu. Ustinova – Doctor of Medical Sciences, Associate Professor, Head of Human Ecology and Life Safety Department (e-mail: ustinova@fcrisk.ru; tel.: +7 (342) 237-25-34; ORCID: <http://orcid.org/0000-0002-9916-5491>).

gaseous matter (volatile organic compounds, ozone, nitrogen oxide, carbon monoxide, sulfur dioxide, etc.). PM in their turn can have various sizes and are usually divided into three basic groups: coarse particles (PM₁₀, diameter is from 2.5 to 10 µm), fine particles (PM_{2.5}, diameter is less than 2.5 µm), and ultrafine particles (diameter is less than 0.1 µm); besides, particulate matter are divided into primary and secondary ones. Primary particulate matter are emitted into the atmosphere directly from a polluting source (dust, metals, soot, salt particles, and plants spores) whereas secondary ones are created from aerosol pollutants through nucleation and gaseous fraction transforming into solid one [4]. Chemical structure of PM varies depending on a polluting source and can include complex organic molecules (formaldehyde, acrolein, benzene, xylene, butadiene, and polycyclic aromatic hydrocarbons), carbon, metals, sulfates, and nitrates [1, 4–7]. The American Heart Association and the European Society of Cardiology determine PM_{2.5} as a risk factor that can cause cardiovascular diseases [8]. Several research works confirmed that exposure to PM_{2.5} created elevated risks of cardiovascular mortality and non-fatal cardiovascular events [3, 9–11]. However, no chemical components or PM_{2.5} sources have been detected to possess any unique toxic effects or peculiarities related to pathogenesis of cardiovascular disorders [3]. It was shown in some research works that frequency of cardiovascular mortality grew by 1 % per each 10 µg/m³ of growing PM_{2.5} concentration in ambient air under short-term exposure [9]. According to meta-analysis data obtained by G. Cesaroni with colleagues, long-term exposure to PM_{2.5} per each 10 µg/m³ of its contents in ambient air resulted in a 6 % growth in mortality due to all reasons (95 % CI 4–8 %) and to an 11 % growth in cardiovascular mortality (95 % CI 5–16 %) [12]. R. Chen with colleagues showed in their study that there was no “safe” bottom limit in a curve showing “dose – effect” dependence for influence exerted by air pollution with PM_{2.5} on cardiovascular prevalence and mortality [13].

H. Mustafic with colleagues showed in their meta-analysis that short-term exposure to aerogenic pollutants (PM_{2.5}, nitrogen, carbon, and sulfur oxides) was associated with a 1–5 % growth in risks of myocardial infarction during the next day [14]. Sub-analysis of acute coronary events in eleven cohorts included into ESCAPE project (more than 100 thousand participants) revealed a statistically significant association between their development and exposure to PM_{2.5} even in concentration that was lower than permitted in the European Union (risk went up by 12 % per each 10 µg/m³ of growing PM₁₀ concentration and by 13 % per each 5 µg/m³ of growing PM_{2.5} concentration) [12]. 6.2 million strokes were analyzed in 28 countries and this allowed establishing rather small (approximately 1 %) but statistically significant increase in frequency of strokes and mortality due to them per each 10 µg/m³ of growing PM_{2.5} concentration [15]. At the same time, according to Woman's Health Initiative and ESCAPE surveys, a risk of stroke grew by 17.5 and 19 % in the USA and Europe accordingly per each 5 µg/m³ of growing PM_{2.5} concentration in ambient air [16, 17]. According to P. Zhang and others, combined exposure to PM₁₀ and nitrogen oxide during 12 years was associated with elevated mortality due to cerebrovascular diseases in China and Korea [18, 19]. Results obtained in The Women's Health Initiative study revealed that long-term aerogenic exposure to PM_{2.5} resulted in a 35 % growth in risks of stroke [17]. Short-term increase in concentrations of gaseous pollutants and PM_{2.5} in ambient air was associated with a growing risk of admission to hospital due to heart failure and related mortality. Each 10 µg/m³ of growing PM_{2.5} concentration led to 1–1.5 % growth in a number of admissions to hospitals due to heart failure [4, 20]. A national cohort study performed in England revealed high frequency of heart failure cases associated with long-term exposure to elevated PM and nitrogen oxide concentrations in ambient air [17–22]; attributive risks of

mortality due to progressing heart failure associated with exposure to aerogenic pollutants could reach 30 % [4, 23].

Pathogenetic mechanisms of cardiovascular pathology associated with exposure to aerogenic pollutants. Multiple studies that have been performed over the last two decades have revealed basic pathogenetic mechanisms of cardiovascular pathology development under exposure to aerogenic pollutants. When aerogenic pollutants penetrate the airways, they start interacting with the mucosa in the upper and lower airways. Coarse and fine particles (PM_{2.5} and PM₁₀) are distributed in the airways in a different way; and since they have different chemical structure, they produce different biologic effects [24]. A primary affecting factor here is PM and other chemicals depositing in airways tissues; it activates three basic mechanisms [25]:

- pro-inflammatory and oxidant mediators are generated and released into blood flow;
- imbalance develops in the autonomic nervous system with prevailing activity of the sympathetic nervous system and disrupted heart rate variability (HRV);
- pollutants penetrate blood flow directly from the lung tissue and this produces direct toxic effects.

Generation of pro-inflammatory and oxidative mediators. Introduction of PM and other chemical components, especially those with redox potential, induces a cascade of oxidative stress reactions and inflammation in lung tissues [9, 22]. Free radicals can both occur in the lung tissue directly due to impacts exerted by some substances, metals, for example, and due to activation of the immune system cells as a result of contacts with PM. Stimulation of redox mechanism with free radicals formation (superoxide and hydroxyl) results in elevated activity of polymorphonuclear leukocytes and production of pro-inflammatory cytokines (interleukin 1 β , interleukin-6, interleukin-8, tumor necrosis factor alpha) and chemokines to achieve sequestration of PM that penetrated the lung tissues. It should be noted that the smaller is a diameter

of particles that have penetrated the lung tissues, the more apparent oxidative stress is induced by them. Several cohort studies revealed there was a relation between exposure to PM (both in occupational groups and population in general) with elevated contents of protein, lipid, and DNA oxidation products in blood and urine. People with already elevated contents of oxidative stress and inflammation products are more susceptible to effects produced by aerogenic pollutants [23, 26, 27]. Besides, PM contacts with the airways mucosa lead to damage to the surfactant system and induce mechanisms of inborn immunity thus creating a focus of chronic low level non-infectious inflammation in the lung tissues. A key hypothesis regarding this mechanism is that the process is deadaptation in its essence and occurring pro-inflammatory and oxidative reactions are not limited to the lung tissues but also result in excessive release of inflammation mediators in peripheral blood flow; thus, a systemic inflammatory reaction develops together with vascular oxidative stress and inflammation as well as endothelial dysfunction [4]. As redox and pro-inflammatory processes develop and generalize, blood cells (polymorphonuclear leukocytes, T-lymphocytes, macrophages, and thrombocytes) become activated; the same goes for a whole cascade of mediators that affect the cardiovascular system including adipocytokines (resistin and adiponectin) and acute-phase proteins (C-reactive protein, fibrinogen, and coagulation factors). These mediators and proteins are activated from adipocytes and hepatocytes accordingly. The aforementioned mechanism induces several basic chronic pathologic processes including atherosclerosis, endothelial dysfunction, cardiac hypertrophy, vasoconstriction, and pro-coagulant changes, and metabolic disorders [28]. However, this mechanism can potentially cause acute exacerbations such as disrupted stability of atherosclerotic plaque with developing acute ischemic syndromes [25].

Developing imbalance in the autonomic nervous system. Several nerve receptors (no-

ciceptive and adrenergic ones) in the lung tissue are activated by PM and other chemicals that penetrate the airways due to breathing; this creates pathological autonomic reflex arcs. This mechanism induces both dysfunction of the central nervous system (activated vegetative centers) and peripheral disorders that become apparent through sympathetic impacts prevailing over parasympathetic ones (changes in heart rate, blood pressure, lower heart rate variability, and disrupted cardiac repolarization) [29]. Activation of the autonomic nervous system and synthesis of vasoactive mediators stimulate so called “neurogenic inflammation” with activated T-lymphocytes, adhesion molecules, pro-inflammatory cytokines, and free radicals [23]. These data indicate there is a close connection between the first and the second mechanism of pathologic influence by aerogenic pollutants and they produce complex effects on the cardiovascular system. Acute reactions induced by this mechanism include predominantly heart rhythm disturbances; chronic ones are arterial hypertension (AH), cardiac and artery hypertrophy, and metabolic disorders [25].

Introduction of pollutants from the lung tissue into blood flow. A few studies have revealed that fine particulate matter as well as soluble chemicals can penetrate blood flow and produce direct adverse effects on vessel walls and blood cells (thrombocytes in particular) and disrupt vasomotor regulation. However, since there are scarce scientific data on the matter, an insight into this mechanism is still considered to be rather vague and contradictory [25, 30].

In scientific literature, the aforementioned mechanisms predominantly explain most typical pathological processes in the cardiovascular system associated with exposure to aerogenic pollutants (AH, atherogenesis, heart rhythm disturbances, resistance to insulin, metabolic syndrome, and type 2 diabetes mellitus). Next we are going to pay closer attention to statistical, pathophysiological and clinical aspects regarding these disorders in relation with exposure to aerogenic pollutants.

Most common diseases, syndromes, and pathological processes that develop under exposure to aerogenic pollutants. Arterial hypertension. A lot of studies concentrate on examining correlation between AH and exposure to aerogenic pollutants. According to Global Burden of Disease Study, elevated blood pressure causes approximately 10.4 million deaths and 208.1 million disability-adjusted life years (DALY) worldwide [2]. An increase by 20 mm Hg in systolic blood pressure (SBP) or by 10 mm Hg in diastolic blood pressure (DBP) in people aged 40–69 is associated with more than 2-time growth in a risk of death due to cardiovascular events [31]. Even a slight decrease in blood pressure leads to significantly improved forecasts; a decrease by 2 mm Hg in SBP results in 5 % drop in mortality due to stroke; 4 % drop, due to ischemic heart diseases (IHD); 3 % drop in overall mortality [32]. Keeping in mind statistical data on cardiovascular prevalence and mortality, we can assume that even slight influence on AH risk factors can significantly improve forecasts for an overall population. A substantial database has been accumulated by now on impacts exerted by ambient air pollution on AH development or induction [8, 10, 11, 33–36]. Research works concentrated on describing influence by aerogenic pollutants on blood pressure, development of persistent AH, and a rise in appealability for emergency aid. According to E. Braunwald, blood pressure grows by 1–4 mm Hg per each $10 \mu\text{g}/\text{m}^3$ of growing $\text{PM}_{2.5}$ concentration in ambient air under short-term exposure. Blood pressure can grow by 8–9 mm Hg during 2–5 days after exposure to elevated $\text{PM}_{2.5}$ concentrations [4, 35]. R. Liang with colleagues performed meta-analysis of 25 studies to show that blood pressure increased by 1.4 / 0.9 mm Hg per each $10 \mu\text{g}/\text{m}^3$ of growing $\text{PM}_{2.5}$ concentration in ambient air [37]. P. Giorgini and colleagues examined 2,078 patients with AH and revealed that average $\text{PM}_{2.5}$ concentration in ambient air equal to $12.6 \pm 8.2 \mu\text{g}/\text{m}^3$ during several previous days was associated with statistically significant rise in blood pressure, by 2.1–3.5 /

1.7–1.8 mm Hg per each standard deviation towards growth in $PM_{2.5}$ concentration. This effect was detected in spite of all research participants constantly being treated with up-to-date hypotensive therapy and ambient air quality was quite optimal if taken in remote prospects [38]. R.D. Brook and others assessed impacts exerted by ambient air pollution on blood pressure of 65 people living in Beijing where $PM_{2.5}$ concentration varied within 9.0–552 $\mu\text{g}/\text{m}^3$. Exposure to high $PM_{2.5}$ concentrations during 1–7 previous days resulted in growth in SBP varying from 2 (95 % CI 0.3–3.7) to 2.7 (95 % CI 0.6–4.8) mm Hg per each standard deviation in $PM_{2.5}$ concentration (67.2 $\mu\text{g}/\text{m}^3$) [39]. Long-term exposure (during the whole previous year) resulted in more apparent hypertensive effects, growth by 7.3 / 9.5 mm Hg [36]. It was established in several research works that “acute” exposure to PM with a wide range of diameters ($PM_{0.1-10}$) and concentrations in ambient air led to rapid rise in blood pressure from 2 to 10 mm Hg and it remained high for several hours [35]. Long-term exposure to $PM_{2.5}$ (during several years) is associated with persistent AH occurrence. An increase in $PM_{2.5}$ concentration in ambient air by 10 $\mu\text{g}/\text{m}^3$ was associated with 13 % of AH prevalence (HR 1.13 (95 % CI 1.05–1.22)) among a cohort made up of 35,303 adult people living on a relatively ecologically clean territory in Canada [40]. Increased appealability for emergency aid due to exacerbated AH was detected during periods when there was growth in ambient air pollution both in countries with high pollution (China) and with lower one (Canada) [41]. Long-term exposure to $PM_{2.5}$ was proven to be associated with an increase in AH-related mortality [42].

Endothelial dysfunction and vasoconstriction. Several studies revealed that exposure to aerogenic pollutants resulted in vasoconstriction and endothelial dysfunction [25, 29]. In particular, S.D. Adar with colleagues established in Multi-Ethnic Study of Atherosclerosis (MESA) performed on a cohort made up of 4,607 patients that exposure to $PM_{2.5}$ influenced reduction of retinal artery

lumen. This effects were also observed under both short-term (during a day) and long-term (2 years, on average) exposure [43]. A similar effect was revealed in a study accomplished in Belgium (84 patients) that concentrated on influence exerted by PM_{10} on retinal artery lumen under short-term exposure [44].

Endothelial dysfunction induced by affecting factors (aerogenic pollutants in particular) is among the most significant mechanisms of cardiovascular pathology. A slight increase in $PM_{2.5}$ concentration in ambient air results in authentic decrease in flow-dependent vasodilatation as it was shown in large cohort studies performed in the USA [45, 46]. An increase in average annual $PM_{2.5}$ concentration in ambient air by 3 $\mu\text{g}/\text{m}^3$ can be compared to effects produced on endothelial function by smoking (a decrease by 0.3 %) or additional 5 years of age [46].

Heart rhythm disturbances. According to literature data, exposure to aerogenic pollutants correlates with electrical instability of the cardiac muscle, changes in heart rate, and lower heart rate variability (HRV) [47–50]. High levels of systemic inflammation and oxidative stress markers were established to predispose to more apparent heart rhythm disturbances under exposure to aerogenic pollutants [49]. However, ventricular arrhythmia and atrial fibrillation induced by exposure to pollutants in most cases were detected in people who already suffered from heart diseases [51]. Moreover, studies that involved controlled inhalation exposure to pollutants didn’t reveal any developing arrhythmia in healthy participants under acute exposure [52].

Changes in HRV were among first biological effects detected under exposure to aerogenic pollutants [25]. Meta-analysis of 29 epidemiologic studies performed on 18,667 patients revealed that an increase in $PM_{2.5}$ concentration in ambient air by 10 $\mu\text{g}/\text{m}^3$ led to statistically significant decrease in both time and frequency HRV parameters and imbalance in the autonomic nervous system towards prevailing sympathetic activity [53]. A decrease

in HRV is a well-known factor that creates risks of cardiovascular death. Several studies revealed disrupted repolarization in ECG under exposure to PM. These data indicate there are certain changes in ion channels of cardiomyocytes and this can induce heart rhythm disturbances in predisposed people, right up to fatal ventricular arrhythmia [54, 55]. Neurogenic effects produced by exposure to aerogenic pollutants on developing imbalance in the autonomic nervous system were confirmed in a study accomplished by C.M. Barbosa in Brazil. When sugar cane was being burnt, PM concentrations grew in ambient air and this resulted in elevated blood pressure in healthy workers due to growing activity of peripheral sympathetic nerves established with microneurography [56].

Atherosclerosis. Long-term exposure to aerogenic pollutants makes for atherosclerosis development [25]. Several research works have concentrated on this effect by pollutants over the last two decades. According to E.H. Wilker long-term exposure to fine-disperse PM in high concentrations is associated with growing carotid intima-media thickness (CIMT), by 1.1 % annually per each $0.26 \mu\text{g}/\text{m}^3$ of growing PM concentration in ambient air [57]. The most significant results were obtained in MESA survey. A series of ultrasound examinations were performed on 5,362 people living in 6 cities in the USA where ambient air pollution was high; they revealed that CIMT grew by $5 \mu\text{m}$ (95 % CI 2.6–7.4) annually per each $2.5 \mu\text{g}/\text{m}^3$ of growing $\text{PM}_{2.5}$ concentration. At the same time, a fall in aerogenic exposure to $\text{PM}_{2.5}$ by $1 \mu\text{g}/\text{m}^3$ resulted in negative CIMT progression ($-2.8 \mu\text{m}$ annually (95 % CI from -1.6 to -3.9)) [58]. First data on influence exerted by ambient air pollution on atherosclerosis progression were obtained by N. Kunzli in a study performed among people in Los Angeles in 2005. Overall, 798 people were examined; they were older than 40 and didn't have either cardiovascular pathology or diabetes in their case history. It was established that carotid intima-media thickness grew by 5.9 % (95 % CI

1.0–10.9) per each $10 \mu\text{g}/\text{m}^3$ of growing $\text{PM}_{2.5}$ concentration in ambient air (within 5.2–26.9 $\mu\text{g}/\text{m}^3$ range) [59]. M. Bauer with colleagues revealed in their study accomplished in 2010 in Germany that a growth in $\text{PM}_{2.5}$ concentration in ambient air by $4.2 \mu\text{g}/\text{m}^3$, and PM_{10} by $6.7 \mu\text{g}/\text{m}^3$, was associated with an increase in CIMT by 4.3 % (95 % CI 1.9–6.7 %) and 1.7 % (-0.7 –4.1 %) accordingly [60]. According to C. Tonne and others, CIMT grows by 5 % (95 % CI 1.9–8.3 %) per each $5.2 \mu\text{g}/\text{m}^3$ of growing PM_{10} concentration [2].

Insulin resistance, type 2 diabetes mellitus, metabolic syndrome. Over the last two decades a lot of data have been accumulated on impacts exerted by aerogenic pollutants on developing diseases and conditions related to insulin resistance (metabolic syndrome, type 2 diabetes mellitus (DM2), and non-alcoholic fatty liver disease) [61, 62]. These effects produced by aerogenic pollutants have been detected both in regions where ambient air pollution is relatively low and in those where it is considerably high [63, 64]. Insulin resistance is a key pathogenetic mechanism responsible for these disorders. As it was shown by R.D. Brook, when 25 healthy adults were exposed to elevated $\text{PM}_{2.5}$ concentrations during 5 days (south-eastern Michigan), it led to a decrease in sensitivity to insulin according to data by Homeostasis Model Assessment of Insulin Resistance Values (HOMA-IR) [65]. The research established that heart rate variability disorders correlated with growing insulin resistance. E.H. Wilker revealed growing insulin resistance and deteriorating control over DM2 in a study conducted in Germany [57]. Long-term exposure to $\text{PM}_{2.5}$ leads to elevated risks of manifest DM2 occurrence simultaneously with growing risks of AH development [61]. H. Chen with colleagues examined 62,012 people living in Canada in their research and showed that DM2 risk grew by 11 % per each $10 \mu\text{g}/\text{m}^3$ of long-term growing $\text{PM}_{2.5}$ concentration in ambient air [63]. F. Liang and others observed 88,397 people in China with 6,439

new DN2 cases detected among them; the authors revealed that exposure to PM_{2.5} increased DM2 risks by 15.7 % (95 % CI 6.42–25.70) per each 10 µg/m³ [66]. A study by Lao X.Q. also established an increase in DM2 risks under long-term exposure to PM_{2.5}. As opposed to the first quartile of PM_{2.5} concentration in ambient air, HR amounted to 1.28 (95 % CI 1.18–1.39), 1.27 (95 % CI 1.17–1.38), and 1.16 (95 % CI 1.07–1.26) for the second, third, and fourth quartiles accordingly [67]. Meta-analysis by I.C. Eze and a review by X. Rao confirmed that DM2 risks grew by 8–13 % per each 10 µg/m³ of growing PM_{2.5} concentration in ambient air [68, 69]. Oxidative stress and chronic metaflammation play a leading role in pathogenesis of metabolic disorders and DM2 associated with exposure to aerogenic pollutants due to activation of pro-inflammatory cytokines and inborn immunity cells in visceral fat depots [61]. This mechanism is quite similar to pathogenesis associated with diabetogenic effects produced by some foods where inflammation plays a leading role in DM2 pathogenesis [70]. Exposure to aerogenic pollutants is associated with elevated levels of tumor necrosis factor, interleukin-6, resistin, and leptin in blood. There is also a growth in concentrations of pro-thrombotic adipokines (plasminogen activator inhibitor-1) and circulating adhesion molecules (ICAM-1, E-selectin). The latter make for leukocytes adhesion to endothelium of post-capillary venules [71]. Imbalance in the autonomic nervous system towards increasing activity of the parasympathetic section also makes a significant contribution to developing insulin resistance. Some lung receptors such as transient receptor potential ankyrin 1 (TRPA1) can be stimulated by aerogenic pollutants and the sympathetic section in the autonomic nervous system through central mechanisms [72]. Endothelial dysfunction often precedes insulin resistance and is associated with disrupted peripheral glucose utilization [73]. Toll-like receptors (TLRs) and nucleotide-binding oligomerization domain-like receptors (NLRs) determine a pathoge-

netic mechanism responsible for a relation between exposure to aerogenic pollutants and obesity / DM2 [74, 75]. Redox reaction products (palmitoyl-arachidonoyl phosphocholine, hyaluronic acid fragments) activate TLR4 and also make for release of chemokines ligand (CCL-2) that activates monocytes [76]. Overall, there are 4 basic mechanisms of immune activation that lead to developing insulin resistance / DM2:

- direct inflammatory / oxidative stress that stimulates alveolar macrophages to synthesize interleukin-1, tumor necrosis factor alpha, interleukin-6, and chemokines (CCL-2 and CCL-5) which determine cellular response in the bone marrow and spleen [77];

- macrophages capturing inhaled pollutants and presenting them to T-lymphocytes in secondary lymphoid organs [78];

- some pollutants (organic compounds, metals) directly penetrating the blood flow with developing vascular inflammation and insulin resistance [79];

- central inflammation mechanisms being involved through reflex arcs connecting receptors in the lung tissue with the brain [73].

Cardio-metabolic effects produced by persistent organic pollutants. Cardio-metabolic effects produced by persistent organic pollutants (POPs) have been described in scientific literature in much lesser volumes than effects by PM. Such compounds are represented by a lot of toxicants such as polychlorinated biphenyls, dioxins, aromatic compounds, and halogen-substituted aromatic hydrocarbons. Biologic effects produced by such compounds on the cardiovascular system are predominantly examined in cross-sectional population and prospect studies that provide an opportunity to trace remote effects produced by low doses of organic toxicants [34]. As a rule, pathogenetic mechanisms of effects by POPs regarding metabolic disorders have few specific features and are just typical pathological processes. Thus, dioxins and polychlorinated biphenyls affect aryl hydrocarbon receptor (AHR). When this receptor is activated, this leads to growing activity

of cytochrome P450 CYP1A1 and formation of reactive oxygen species together with developing low level inflammation [80]. Simultaneously apoptosis is disrupted and changes occur in the cellular lifecycle, there is also lipid oxidation and accelerated atherosclerotic processes in vascular walls [81]. Some substances influence peroxisome proliferator-activated receptors (PPAR) that disrupt adipocytes differentiation, lipid metabolism, reduce sensitivity to insulin and hence increase risks of DM2 development [82].

There are data in scientific literature on effects produced by POPs on risk factors causing cardiovascular diseases (AH, obesity, and DM2) and ultimate events (myocardial infarction, stroke, or diseases of peripheral arteries). There was an incident in Yucheng that involved mass exposure to polychlorinated biphenyls and polychlorinated dibenzofurans; as a result, a risk of AH development grew by more than 3 times over the next 24 years among women with chloracne manifestations in comparison with patients without this pathology [83]. A.V. Sergeev performed a study among people living on a territory polluted with POPs and revealed that AH risks was by 19 % higher for them (95 % CI 9–31 %) in comparison with the reference group [84]. NHANES cross-sectional study revealed that relative AH risk amounted to 1.8 (95 % CI 1.2–2.7) in the highest quartile as per environmental pollution with polychlorinated biphenyls [85]. It was established in the same study performed with 524 people participating in it that concentrations of dioxin and polychlorinated dibenzofurans in blood correlated with risks of developing AH in women, relative risks being equal to 5–6 for the highest quartile against the lowest one [86]. A cross-sectional study that involved 758 participants (Anniston, the USA) who lived on a territory polluted with polychlorinated biphenyls revealed that relative risk of developing AH amounted to 4.1 (95 % CI 1.3–14) for the upper tertile in comparison with the lower one [34].

Effects produced by POPs are associated with developing metabolic syndrome (MS). At

present there are only results of cross sectional studies that focused on influence by POPs on MS occurrence. A national study was performed in Japan and as a result it was established that contents of dioxins and polychlorinated biphenyls in blood of more than 1,300 people correlated with developing MS (*OR* 3.2–4.8 when the upper and the lower quartiles were compared). Elevated POPs concentration in blood was associated with greater frequency of all components in metabolic syndrome [87]. NHANES survey revealed that pesticides concentration in plasma correlated authentically with developing MS (*OR* 5.3, 95 % CI 2.5–11, when the upper and the lower quartiles were compared). Besides, elevated pesticides contents in blood (the upper quartile against the lower one) were associated with *OR* = 2.4 for the waist circumference; 7.1, for triglycerides; 2.3, for low density lipoproteins; 5.6, for glucose; 1.8, for AH [88]. S.K. Park with colleagues (Korea) compared 50 patients with MS with a reference group and revealed a correlation between MS and concentrations of pesticides, beta-hexachlorocyclohexane and heptachlor epoxide in blood plasma (*OR* 4.4–6.0 for the upper and lower quartiles) [89]. Insulin resistance under exposure to POPs was examined as a key pathogenetic mechanism of developing MS in NHANES survey in 749 patients without DM2. 19 various POPs were examined; the most apparent association with HOMA-IR index showing insulin resistance was revealed for pesticides (*OR* = 3.8 for the upper quartile) and this dependence was maximum among people with large waist circumference [90]. PIVUS survey focused on examining an association between POPs and fat mass using DXA. Low-chlorinated polychlorinated biphenyls turned out to correlate positively with fat mass whereas high-chlorinated ones had an inverse correlation with the parameter [91]. This difference in influence exerted by POPs chlorination can be due to pharmacokinetic properties and the fact that low-chlorinated compounds have a shorter semiejection period. A peak in using such POPs was in 70ties last century. M.S. Wolff with colleagues noted that a correlation be-

tween POPs and overweight was always negative under short-term exposure to toxicants due to their depositing in fat tissue. However, it became positive after 2 or 3 semiejction periods in case there was no further POPs introduction [92]. NHANES survey revealed a direct correlation between dioxin concentrations in blood and body mass index (BMI) both in men and women [93]. A population study (13,000 participants) accomplished in Japan allowed establishing positive dependence between concentration of polychlorinated biphenyls in blood and BMI [87]. CARDIA survey yielded similar results during 25 years of observation [94]. Several studies revealed elevated DM2 risks under long-term exposure to POPs. O. Vasiliu with colleagues established in their study (more than 1,300 participants, Michigan) that odds ratio for developing DM2 amounted to 2.0–3.0 for women and 1.7 for men depending on concentration of polybrominated biphenyls in blood (the upper quartile compared with the lower one) [95]. It was established in CARDIA survey that elevated concentrations of trans-nona-chlodane pesticide and some polychlorinated biphenyls in blood were predictors of developing DM2. The authors noted that the effect occurred even if a rise in concentration of trans-nona-chlodane in blood was relatively low ($OR = 5.3$ for the second quartile against the first one) [94].

Long-term exposure to POPs can lead to developing atherosclerotic processes. PIVUS cross-sectional study established that polychlorinated biphenyls influenced formation of atherosclerotic plaques even after statistical adjustment per 10 well-known risk factors, lipids included. A similar effect was revealed for phthalate metabolites [96]. A study by IARC that involved 21,863 workers revealed that long-term exposure to dioxins at a workplace was associated with developing IHD (RR 1.6, 95 % CI 1.2–2.2). Relative risk of stroke amounted to 1.5 (95 % CI 0.8–2.8) in the same cohort [97]. A.V. Sergeev and I. Shcherbatykh examined a population living on a territory polluted with POPs (New York) and revealed that relative risks of myocardial infarction grew by

20 % (95 % CI 3–39 %); stroke, by 10 % (95 % CI 1.0–1.2) [84, 98]. NHANES survey established that relative risk of myocardial infarction under exposure to bisphenol A amounted to 1.2 (95 % CI 1.1–1.4) per one standard deviation in bisphenol A concentration in urine [99].

Effects produced by prevention activities aimed at reducing influence by ambient air pollution on the cardiovascular system. Randomized studies showed there was a direct protective effect produced by a decrease in pollutants concentrations in ambient air. Use of filtration devices and face masks that are able to filter PM results in decreasing blood pressure, microvascular function improvement, and lower levels of inflammatory biomarkers in adults exposed to $PM_{2.5}$ [100–102]. Results obtained by C.A. Pope with colleagues revealed that in 1970–2000 average life expectancy in the USA grew by 0.61 year per each $10 \mu\text{g}/\text{m}^3$ of declining $PM_{2.5}$ concentration in ambient air (demographic, socioeconomic, and behavioral factors taken into account) [103]. According to M. Morishita, declining $PM_{2.5}$ concentration in ambient air in 1970–1990ties resulted in overall mortality going down by 27 %, and cardiovascular mortality, by 31 %, per each $10 \mu\text{g}/\text{m}^3$ [104].

Activities aimed at reducing ambient air pollution may be implemented over decades; however, even a short-term decrease in pollutants concentration in ambient air (as it was shown during the Olympics in Beijing) leads to a rapid fall in levels of inflammation markers, oxidative stress, and thrombosis [105]. According to data provided by the US Environmental Protection Agency, activities aimed at ambient air purification prevented more than 160,000 deaths and 130,00 myocardial infarctions in 2010. It is pointed out that activities aimed at reducing concentrations of pollutants are likely to have more apparent effects in countries where quality of ambient air is rather poor [23].

Funding. The research was not granted any financial support.

Conflict of interests. The authors declare there is no any conflict of interests.

References

1. Hadley M.B., Baumgartner J., Vedanthan R. Developing a Clinical Approach to Mitigating Risks of Air Pollution and Protecting Cardiovascular Health *Circulation*, 2018, vol. 137, no. 7, pp. 725–742. DOI: 10.1161/CIRCULATIONAHA.117.030377
2. GBD 2016 Risk Factors Collaborators. Global, regional, and national comparative risk assessment of 84 behavioural, environmental and occupational, and metabolic risks or clusters of risks, 1990–2016: a systematic analysis for the Global Burden of Disease Study 2016. *Lancet*, 2017, vol. 390, no. 10100, pp. 1345–1422. DOI: 10.1016/S0140-6736(17)32366-8
3. Cohen A.J., Brauer M., Burnett R., Anderson H.R., Frostad J., Estep K., Balakrishnan K., Brunekreef B. [et al.]. Estimates and 25-year trends of the global burden of disease attributable to ambient air pollution: an analysis of data from the Global Burden of Diseases Study 2015. *Lancet*, 2017, vol. 389, no. 10082, pp. 1907–1918. DOI: 10.1016/S0140-6736(17)30505-6
4. Zipes D.P., Libby P., Bonow R.O., Mann D.L., Tomaselli G.F. Braunwald's Heart Disease: A Textbook of Cardiovascular Medicine 11th Edition. Elsevier, 2018, 2040 p.
5. Cullen M.R. Invited commentary: the search for preventable causes of cardiovascular disease – whither work? *Am. J. Epidemiol.*, 2009, vol. 169, no. 12, pp. 1422–1425. DOI: 10.1093/aje/kwp078
6. Fang S.C., Cassidy A., Christiani D.C. A systematic review of occupational exposure to particulate matter and cardiovascular disease. *Int. J. Environ. Res. Public Health*, 2010, vol. 7, no. 4, pp. 1773–1806. DOI: 10.3390/ijerph7041773
7. Secrest M.H., Schauer J.J., Carter E.M., Baumgartner J. Particulate matter chemical component concentrations and sources in settings of household solid fuel use. *Indoor Air*, 2017, vol. 27, no. 6, pp. 1052–1066. DOI: 10.1111/ina.12389
8. Newby D.E., Mannucci P.M., Tell G.S., Baccarelli A.A., Brook R.D., Donaldson K., Forastiere F., Franchini M. [et al.]. Expert position paper on air pollution and cardiovascular disease. *Eur. Heart J.*, 2015, vol. 36, no. 2, pp. 83–93b. DOI: 10.1093/eurheartj/ehu458
9. Brook R.D., Rajagopalan S., Pope C.A. 3rd, Brook J.R., Bhatnagar A., Diez-Roux A.V., Holguin F., Hong Y. [et al.]. Particulate matter air pollution and cardiovascular disease: An update to the scientific statement from the American Heart Association. *Circulation*, 2010, vol. 121, no. 21, pp. 2331–2378. DOI: 10.1161/CIR.0b013e3181dbecel
10. Munzel T., Sorensen M., Gori T., Schmidt F.P., Rao X., Brook J., Chen L.C., Brook R.D., Rajagopalan S. Environmental stressors and cardio-metabolic disease: part I – epidemiologic evidence supporting a role for noise and air pollution and effects of mitigation strategies. *Eur. Heart J.*, 2017, vol. 38, no. 8, pp. 550–556. DOI: 10.1093/eurheartj/ehw269
11. Munzel T., Sorensen M., Gori T., Schmidt F.P., Rao X., Brook J., Chen L.C., Brook R.D., Rajagopalan S. Environmental stressors and cardio-metabolic disease: part II – mechanistic insights. *Eur. Heart J.*, 2017, vol. 38, no. 8, pp. 557–564. DOI: 10.1093/eurheartj/ehw294
12. Cesaroni G., Forastiere F., Stafoggia M., Andersen Z.J., Badaloni C., Beelen R., Caracciolo B., de Faire U. [et al.]. Long-term exposure to ambient air pollution and incidence of acute coronary events: prospective cohort study and meta-analysis in 11 European cohorts from the ESCAPE Project. *Br. Med. J.*, 2014, vol. 348, pp. f7412. DOI: 10.1136/bmj.f7412
13. Chen R., Kan H., Chen B., Huang W., Bai Z., Song G., Pan G., CAPES Collaborative Group. Association of particulate air pollution with daily mortality: the China Air Pollution and Health Effects Study. *Am. J. Epidemiol.*, 2012, vol. 175, no. 11, pp. 1173–1181. DOI: 10.1093/aje/kwr425
14. Mustafic H., Jabre P., Caussin C., Murad M.H., Escolano S., Tafflet M., Perier M.C., Marjion E. [et al.]. Main air pollutants and myocardial infarction: a systematic review and meta-analysis. *J. Am. Med. Assoc.*, 2012, vol. 307, no. 7, pp. 713–721. DOI: 10.1001/jama.2012.126
15. Yang W.S., Wang X., Deng Q., Fan W.Y., Wang W.Y. An evidence-based appraisal of global association between air pollution and risk of stroke. *Int. J. Cardiol.*, 2014, vol. 175, no. 2, pp. 307–313. DOI: 10.1016/j.ijcard.2014.05.044
16. Stafoggia M., Cesaroni G., Peters A., Andersen Z.J., Badaloni C., Beelen R., Caracciolo B., Cyrus J. [et al.]. Long-term exposure to ambient air pollution and incidence of cerebrovascular events: results from eleven European cohorts within the ESCAPE project. *Environ. Health Perspect.*, 2014, vol. 122, no. 9, pp. 919–925. DOI: 10.1289/ehp.1307301
17. Miller K.A., Siscovick D.S., Sheppard L., Shepherd K., Sullivan J.H., Anderson G.L., Kaufman J.D. Long-term exposure to air pollution and incidence of cardiovascular events in women. *N. Engl. J. Med.*, 2007, vol. 356, no. 5, pp. 447–458. DOI: 10.1056/NEJMoa054409
18. Zhang P., Dong G., Sun B., Zhang L., Chen X., Ma N., Yu F., Guo H. Long-term exposure to ambient air pollution and mortality due to cardiovascular disease and cerebrovascular disease in Shenyang, China. *PLoS One*, 2011, vol. 6, no. 6, pp. e20827. DOI: 10.1371/journal.pone.0020827
19. Hong Y.C., Lee J.T., Kim H., Ha E.H., Schwartz J., Christiani D.C. Effects of air pollutants on acute stroke mortality. *Environ. Health Perspect.*, 2002, vol. 110, no. 2, pp. 187–191. DOI: 10.1289/ehp.02110187
20. Shah A.S.V., Langrish J.P., Nair H., McAllister D.A., Hunter A.L., Donaldson K., Newby D.E., Mills N.L. Global association of air pollution and heart failure: a systematic review and meta-analysis. *Lancet*, 2013, vol. 382, no. 9897, pp. 1039–1048. DOI: 10.1016/S0140-6736(13)60898-3
21. Atkinson R.W., Carey I.M., Kent A.J., van Staa T.P., Anderson H.R., Cook D.G. Long-term exposure to outdoor air pollution and incidence of cardiovascular diseases. *Epidemiology*, 2013, vol. 24, no. 1, pp. 44–53. DOI: 10.1097/EDE.0b013e318276ccb8
22. Kelly F., Fussell J.C. Role of oxidative stress in cardiovascular disease outcomes following exposure to ambient air pollution. *Free Radic. Biol. Med.*, 2017, vol. 110, pp. 345–367. DOI: 10.1016/j.freeradbiomed.2017.06.019
23. Cosselman K.E., Navas-Acien A., Kaufman J.D. Environmental factors in cardiovascular disease. *Nat. Rev. Cardiol.*, 2015, vol. 12, no. 11, pp. 627–642. DOI: 10.1038/nrcardio.2015.152
24. Falcon-Rodriguez C.I., Osornio-Vargas A.R., Sada-Ovalle I., Segura-Medina P. Aeroparticles, composition, and lung diseases. *Front. Immunol.*, 2016, vol. 7, pp. 3. DOI: 10.3389/fimmu.2016.00003
25. Franklin B.A., Brook R., Pope C.A. 3rd. Air pollution and cardiovascular disease. *Curr. Probl. Cardiol.*, 2015, vol. 40, no. 5, pp. 207–238. DOI: 10.1016/j.cpcardiol.2015.01.003
26. Langrish J.P., Unosson J., Bosson J., Barath S., Muala A., Blackwell S., Söderberg S., Pourazar J. [et al.]. Altered nitric oxide bioavailability contributes to diesel exhaust inhalation-induced cardiovascular dysfunction in man. *J. Am. Heart Assoc.*, 2013, vol. 2, no. 1, pp. e004309. DOI: 10.1161/JAHA.112.004309
27. Gandhi S.K., Rich D.Q., Ohman-Strickland P.A., Kipen H.M., Gow M. Plasma nitrite is an indicator of acute changes in ambient air pollutant concentrations. *Inhal. Toxicol.*, 2014, vol. 26, no. 7, pp. 426–434. DOI: 10.3109/08958378.2014.913216
28. Ruckerl R., Hampel R., Breitner S., Cyrus J., Kraus U., Carter J., Dailey L., Devlin R.B. [et al.]. Associations between ambient air pollution and blood markers of inflammation and coagulation/fibrinolysis in susceptible populations. *Environment international*, 2014, vol. 70, pp. 32–49. DOI: 10.1016/j.envint.2014.05.013
29. Gold D.R., Mittleman M.A. New insights into pollution and the cardiovascular system 2010 to 2012. *Circulation*, 2013, vol. 127, no. 18, pp. 1903–1913. DOI: 10.1161/CIRCULATIONAHA.111.064337
30. Sanidas E.J., Papadopoulos P.D., Grassos H., Velliou R., Tsioufis M., Barbetseas J., Papademetriou V. Air pollution and arterial hypertension. A new risk factor is in the air. *J. Am. Soc. Hypertens.*, 2017, vol. 11, no. 11, pp. 709–715. DOI: 10.1016/j.jash.2017.09.008

31. Lewington S., Clarke R., Qizilbash N., Peto R., Collins R., Prospective Studies Collaboration. Age-specific relevance of usual blood pressure to vascular mortality: A meta-analysis of individual data for one million adults in 61 prospective studies. *Lancet*, 2002, vol. 360, no. 9349, pp. 1903–1913. DOI: 10.1016/s0140-6736(02)11911-8
32. Whelton P.K., He J., Appel L.J., Cutler J.A., Havas S., Kotchen T.A., Roccella E.J., Stout R. [et al.]. Primary prevention of hypertension: Clinical and public health advisory from The National High Blood Pressure Education Program. *Journal of the American Medical Association*, 2002, vol. 288, no. 15, pp. 1882–1888. DOI: 10.1001/jama.288.15.1882
33. Yang B.Y., Qian Z., Howard S.W., Vaughn M.G., Fan S.J., Liu K.K., Dong G.H. Global association between ambient air pollution and blood pressure: A systematic review and meta-analysis. *Environ. Pollut.*, 2018, vol. 235, pp. 576–588. DOI: 10.1016/j.envpol.2018.01.001
34. Lind L., Lind P.M. Can persistent organic pollutants and plastic-associated chemicals cause cardiovascular disease? *Journal of internal medicine*, 2012, vol. 271, no. 6, pp. 537–553. DOI: 10.1111/j.1365-2796.2012.02536.x
35. Giorgini P., Di Giosia P., Grassi D., Rubenire M., Brook R.D., Ferri C. Air pollution exposure and blood pressure: an updated review of the literature. *Curr. Pharm. Des.*, 2015, vol. 22, no. 1, pp. 28–51. DOI: 10.2174/1381612822666151109111712
36. Hypertension: a companion to Braunwald's Heart Disease. 3rd edition. In: G.L. Bakris, M.J. Sorrentino eds. Elsevier, 2017, 520 p.
37. Liang R., Zhang B., Zhao X., Ruan Y., Lian H., Fan Z. Effect of exposure to PM_{2.5} on blood pressure: a systematic review and meta-analysis. *J. Hypertens.*, 2014, vol. 32, no. 11, pp. 2130–2140. DOI: 10.1097/HJH.0000000000000342
38. Giorgini P., Rubenire M., Das R., Gracik T., Wang L., Morishita M., Bard R.L., Jackson E.A. [et al.]. Particulate matter air pollution and ambient temperature: opposing effects on blood pressure in high-risk cardiac patients. *J. Hypertens.*, 2015, vol. 33, no. 10, pp. 2032–2038. DOI: 10.1097/HJH.0000000000000663
39. Brook R.D., Sun Z., Brook J.R., Zhao X., Ruan Y., Yan J., Mukherjee B., Rao X. [et al.]. Extreme air pollution conditions adversely affect blood pressure and insulin resistance: the air pollution and cardiometabolic disease study. *Hypertension*, 2016, vol. 67, no. 1, pp. 77–85. DOI: 10.1161/HYPERTENSIONAHA.115.06237
40. Chen H., Burnett R.T., Kwong J.C., Villeneuve P.J., Goldberg M.S., Brook R.D., van Donkelaar A., Jerrett M. [et al.]. Spatial association between ambient fine particulate matter and incident hypertension. *Circulation*, 2014, vol. 129, no. 5, pp. 562–569. DOI: 10.1161/CIRCULATIONAHA.113.003532
41. Brook R.D., Kousha T. Air pollution and emergency department visits for hypertension in Edmonton and Calgary, Canada: a case-crossover study. *Am. J. Hypertens.*, 2015, vol. 28, no. 9, pp. 1121–1126. DOI: 10.1093/ajh/hpu302
42. Pope C.A. 3rd., Turner M.C., Burnett R.T., Jerrett M., Gapstur S.M., Diver W.R., Krewski D., Brook R.D. Relationships between fine particulate air pollution, cardiometabolic disorders, and cardiovascular mortality. *Circ. Res.*, 2015, vol. 116, no. 1, pp. 108–115. DOI: 10.1161/CIRCRESAHA.116.305060
43. Adar S.D., Klein R., Klein B.E., Szpiro A.A., Cotch M.F., Wong T.Y., O'Neill M.S., Shrager S. [et al.]. Air pollution and the microvasculature: a cross-sectional assessment of in vivo retinal images in the population-based multi-ethnic study of atherosclerosis (MESA). *PLoS Med.*, 2010, vol. 7, no. 11, pp. e1000372. DOI: 10.1371/journal.pmed.1000372
44. Louwies T., Panis L.I., Kicinski M., De Boever P., Nawrot T.S. Retinal microvascular responses to short-term changes in particulate air pollution in healthy adults. *Environ. Health Perspect.*, 2013, vol. 121, no. 9, pp. 1011–1016. DOI: 10.1289/ehp.1205721
45. Wilker E.H., Ljungman P.L., Rice M.B., Kloog I., Schwartz J., Gold D.R., Koutrakis P., Vita J.A. [et al.]. Relation of long-term exposure to air pollution to brachial artery flow-mediated dilation and reactive hyperemia. *Am. J. Cardiol.*, 2014, vol. 113, no. 12, pp. 2057–2063. DOI: 10.1016/j.amjcard.2014.03.048
46. Krishnan R.M., Adar S.D., Szpiro A.A., Jorgensen N.W., Van Hee V.C., Barr R.G., O'Neill M.S., Herrington D.M. [et al.]. Vascular responses to long- and short-term exposure to fine particulate matter: MESA Air (Multi-Ethnic Study of Atherosclerosis and Air Pollution). *J. Am. Coll. Cardiol.*, 2012, vol. 60, no. 21, pp. 2158–2166. DOI: 10.1016/j.jacc.2012.08.973
47. Zanobetti A., Stone P.H., Speizer F.E., Schwartz J.D., Coull B.A., Suh H.H., Nearing B.D., Mittleman M.A. [et al.]. T-wave alternans, air pollution and traffic in high-risk subjects. *Am. J. Cardiol.*, 2009, vol. 104, no. 5, pp. 665–670. DOI: 10.1016/j.amjcard.2009.04.046
48. Cakmak S., Dales R., Kauri L.M., Mahmud M., Van Ryswyk K., Vanos J., Liu L., Kumarathasan P. [et al.]. Metal composition of fine particulate air pollution and acute changes in cardiorespiratory physiology. *Environ. Pollut.*, 2014, vol. 189, pp. 208–214. DOI: 10.1016/j.envpol.2014.03.004
49. Lee M.-S., Eum K.-D., Fang S.C., Rodrigues E.G., Modest G.A., Christiani D.C. Oxidative stress and systemic inflammation as modifiers of cardiac autonomic responses to particulate air pollution. *Int. J. Cardiol.*, 2014, vol. 176, no. 1, pp. 166–170. DOI: 10.1016/j.ijcard.2014.07.012
50. Park S.K., Auchincloss A.H., O'Neill M.S., Prineas R., Correa J.C., Keeler J., Barr R.G., Kaufman J.D., Diez Roux A.V. Particulate air pollution, metabolic syndrome, and heart rate variability: the multi-ethnic study of atherosclerosis (MESA). *Environ. Health Perspect.*, 2010, vol. 118, no. 10, pp. 1406–1411. DOI: 10.1289/ehp.0901778
51. Anderson H.R., Armstrong B., Hajat S., Harrison R., Monk V., Poloniecki J., Timmis A., Wilkinson P. Air pollution and activation of implantable cardioverter defibrillators in London. *Epidemiology*, 2010, vol. 21, no. 3, pp. 405–413. DOI: 10.1097/EDE.0b013e3181d61600
52. Langrish J.P., Watts S.J., Hunter A.J., Shah A.S.V., Bosson J.A., Unosson J., Barath S., Lundbäck M. [et al.]. Controlled exposures to air pollutants and risk of cardiac arrhythmia. *Environ. Health Perspect.*, 2014, vol. 122, no. 7, pp. 747–753. DOI: 10.1289/ehp.1307337
53. Pieters N., Plusquin M., Cox B., Kicinski M., Vanronsveld J., Nawrot T.S. An epidemiological appraisal of the association between heart rate variability and particulate air pollution: a meta-analysis. *Heart*, 2012, vol. 98, no. 15, pp. 1127–1135. DOI: 10.1136/heartjnl-2011-301505
54. Liao D., Shaffer M.L., Rodriguez-Colon S., He F., Li X., Wolbrette D.L., Yanosky J., Cascio W.E. Acute adverse effects of fine particulate air pollution on ventricular repolarization. *Environ. Health Perspect.*, 2010, vol. 118, no. 7, pp. 1010–1015. DOI: 10.1289/ehp.0901648
55. Sivagangabalan G., Spears D., Masse S., Urch B., Brook R.D., Silverman F., Gold D.R., Lukic K.Z. [et al.]. The effect of air pollution on spatial dispersion of myocardial repolarization in healthy human volunteers. *J. Am. Coll. Cardiol.*, 2011, vol. 57, no. 2, pp. 198–206. DOI: 10.1016/j.jacc.2010.08.625
56. Barbosa C.M., Terra-Filho M., de Albuquerque A.L., Di Giorgi D., Grupi C., Negrão C.E., Pinto Brandão Rondon M.U., Martinez D.G. [et al.]. Burnt sugarcane harvesting – cardiovascular effects on a group of healthy workers, Brazil. *PLoS One*, 2012, vol. 7, no. 9, pp. e46142. DOI: 10.1371/journal.pone.0046142
57. Wilker E.H., Mittleman M.A., Coull B.A., Gryparis A., Bots M.L., Schwartz J., Sparrow D. Long-term exposure to black carbon and carotid intima-media thickness: the normative aging study. *Environ. Health Perspect.*, 2013, vol. 121, no. 9, pp. 1061–1067. DOI: 10.1289/ehp.1104845
58. Adar S.D., Sheppard L., Vedal S., Polak J.F., Sampson P.D., Diez Roux A.V., Budoff M.M., Jacobs D.R. Jr. [et al.]. Fine particulate air pollution and the progression of carotid intima-medial thickness: a prospective cohort study from the multi-ethnic study of atherosclerosis and air pollution. *PLoS Med.*, 2013, vol. 10, no. 4, pp. e1001430. DOI: 10.1371/journal.pmed.1001430
59. Kunzli N., Jerrett M., Mack W.J., Beckerman B., La Bree L., Gilliland F., Thomas D., Peters J., Hodis H.N. Ambient air pollution and atherosclerosis in Los Angeles. *Environ. Health Perspect.*, 2005, vol. 113, no. 2, pp. 201–206. DOI: 10.1289/ehp.7523

60. Bauer M., Moebus S., Mohlenkamp S., Dragano N., Nonnemacher M., Fuchsluger M., Kessler C., Jakobs H. [et al.]. Urban particulate matter air pollution is associated with subclinical atherosclerosis: results from the HNR (Heinz Nixdorf Recall) study. *J. Am. Coll. Cardiol.*, 2010, vol. 56, no. 22, pp. 1803–1808. DOI: 10.1016/j.jacc.2010.04.065
61. Rajagopalan S., Brook R.D. Air pollution and type 2 diabetes: mechanistic insights. *Diabetes*, 2012, vol. 61, no. 12, pp. 3037–3045. DOI: 10.2337/db12-0190
62. Brook R.D., Newby D.E., Rajagopalan S. Pollution and Cardiometabolic Disease: An Update and Call for Clinical Trials. *Am. J. Hypertens.*, 2017, vol. 31, no. 1, pp. 1–10. DOI: 10.1093/ajh/hpx109
63. Chen H., Burnett R.T., Kwong J.C., Villeneuve P.J., Goldberg M.S., Brook R.D., van Donkelaar A., Jerrett M. [et al.]. Risk of incident diabetes in relation to long-term exposure to fine particulate matter in Ontario, Canada. *Environ. Health Perspect.*, 2013, vol. 121, no. 7, pp. 804–810. DOI: 10.1289/ehp.1205958
64. Liu C., Yang C., Zhao Y., Ma Z., Bi J., Liu Y., Meng X., Wan Y. [et al.]. Associations between long-term exposure to ambient particulate air pollution and type 2 diabetes prevalence, blood glucose and glycosylated hemoglobin levels in China. *Environ. Int.*, 2016, vol. 92–93, pp. 416–421. DOI: 10.1016/j.envint.2016.03.028
65. Brook R.D., Xu X., Bard L.R., Dvonch J.T., Morishita M., Kaciroti N., Sun Q., Harkema J., Rajagopalan S. Reduced metabolic insulin sensitivity following sub-acute exposures to low levels of ambient fine particulate matter air pollution. *Sci. Total Environ.*, 2013, vol. 448, pp. 66–71. DOI: 10.1016/j.scitotenv.2012.07.034
66. Liang F., Yang X., Liu F., Li J., Xiao Q., Chen J., Liu X., Cao J. [et al.]. Long-term exposure to ambient fine particulate matter and incidence of diabetes in China: A cohort study. *Environment International*, 2019, vol. 126, pp. 568–575. DOI: 10.1016/j.envint.2019.02.069
67. Lao X.Q., Guo Q., Chang L., Bo Y., Zhang Z., Chuang Y.C., Jiang W.K., Lin C. [et al.]. Long-term exposure to ambient fine particulate matter (PM_{2.5}) and incident type 2 diabetes: a longitudinal cohort study. *Diabetologia*, 2019, vol. 62, no. 5, pp. 759–769. DOI: 10.1007/s00125-019-4825-1
68. Eze I.C., Hemkens L.G., Bucher H.C., Hofmann B., Schindler C., Kunzli N., Schikowski T., Probst-Hensch N.M. Association between ambient air pollution and diabetes mellitus in Europe and North America: systematic review and meta-analysis. *Environ. Health Perspect.*, 2015, vol. 123, no. 5, pp. 381–389. DOI: 10.1289/ehp.1307823
69. Rao X., Montresor-Lopez J., Puett R., Rajagopalan S., Brook R.D. Ambient air pollution: an emerging risk factor for diabetes mellitus. *Curr. Diab. Rep.*, 2015, vol. 15, no. 6, pp. 603. DOI: 10.1007/s11892-015-0603-8
70. Shoelson S.E., Lee J., Goldfine A.B. Inflammation and insulin resistance. *J. Clin. Invest.*, 2006, vol. 116, no. 7, pp. 1793–1801. DOI: 10.1172/JCI29069
71. Sun Q., Yue P., Deiluiis J.A., Lumeng C.N., Kampfrath T., Mikolaj M.B., Cai Y., Ostrowski M.C. [et al.]. Ambient air pollution exaggerates adipose inflammation and insulin resistance in a mouse model of diet-induced obesity. *Circulation*, 2009, vol. 119, no. 4, pp. 538–546. DOI: 10.1161/CIRCULATIONAHA.108.799015
72. Simon S.A., Liedtke W. How irritating: the role of TRPA1 in sensing cigarette smoke and aerogenic oxidants in the airways. *J. Clin. Invest.*, 2008, vol. 118, no. 7, pp. 2383–2386. DOI: 10.1172/JCI36111
73. Baron A.D., Steinber H.O., Chaker H., Leaming R., Johnson A., Brechtel G. Insulin-mediated skeletal muscle vasodilation contributes to both insulin sensitivity and responsiveness in lean humans. *J. Clin. Invest.*, 1995, vol. 96, no. 2, pp. 786–792. DOI: 10.1172/JCI118124
74. Shi H., Kokoeva M.V., Inouye K., Tzameli I., Yin H., Flier J.S. TLR4 links innate immunity and fatty acid-induced insulin resistance. *J. Clin. Invest.*, 2006, vol. 116, no. 11, pp. 3015–3025. DOI: 10.1172/JCI28898
75. Vandanmagsar B., Youm Y.H., Ravussin A., Galgani J.E., Stadler K., Mynatt R.L., Ravussin E., Stephens J.M., Dixit V.D. The NLRP3 inflammasome instigates obesity-induced inflammation and insulin resistance. *Nat. Med.*, 2011, vol. 17, no. 2, pp. 179–188. DOI: 10.1038/nm.2279
76. Imai Y., Kuba K., Neely G.G., Yaghubian-Malhami R., Perkmann T., van Loo G., Ermolaeva M., Veldhuizen R. [et al.]. Identification of oxidative stress and Toll-like receptor 4 signaling as a key pathway of acute lung injury. *Cell*, 2008, vol. 133, no. 2, pp. 235–249. DOI: 10.1016/j.cell.2008.02.043
77. Kampfrath T., Maiseyue A., Ying Z., Shah Z., Deiluiis J.A., Xu X., Kherada N., Brook R.D. [et al.]. Chronic fine particulate matter exposure induces systemic vascular dysfunction via NADPH oxidase and TLR4 pathways. *Circ. Res.*, 2011, vol. 108, no. 6, pp. 716–726. DOI: 10.1161/CIRCRESAHA.110.237560
78. Deiluiis J.A., Kampfrath T., Zhong J., Oghumu S., Maiseyue A., Chen L.C., Sun Q., Satoskar A.R., Rajagopalan S. Pulmonary T cell activation in response to chronic particulate air pollution. *Am. J. Physiol. Lung Cell Mol. Physiol.*, 2012, vol. 302, no. 4, pp. L399–409. DOI: 10.1152/ajplung.00261.2011
79. Dominici F., Peng R.D., Ebisu K., Zeger S.L., Samet J.M., Bell M.L. Does the effect of PM₁₀ on mortality depend on PM nickel and vanadium content? A reanalysis of the NMMAPS data. *Environ. Health Perspect.*, 2007, vol. 115, no. 12, pp. 1701–1703. DOI: 10.1289/ehp.10737
80. Shertzer H.G., Nebert D.W., Puga A., Ary M., Sonntag D., Dixon K., Robinson L.J., Cianciolo E., Dalton T.P. Dioxin causes a sustained oxidative stress response in the mouse. *Biochem. Biophys. Res. Commun.*, 1998, vol. 253, no. 1, pp. 44–48. DOI: 10.1006/bbrc.1998.9753
81. Nebert D.W., Roe A.L., Dieter M.Z., Solis W.A., Yang Y., Dalton T.P. Role of the aromatic hydrocarbon receptor and [Ah] gene battery in the oxidative stress response, cell cycle control, and apoptosis. *Biochem. Pharmacol.*, 2000, vol. 59, no. 1, pp. 65–85. DOI: 10.1016/s0006-2952(99)00310-x
82. Desvergne B., Feige J.N., Casals-Casas C. PPAR-mediated activity of phthalates: a link to the obesity epidemic? *Mol. Cell Endocrinol.*, 2009, vol. 304, no. 1–2, pp. 43–48. DOI: 10.1016/j.mce.2009.02.017
83. Wang S.L., Tsai P.C., Yang C.Y., Guo Y.L. Increased risk of diabetes and polychlorinated biphenyls and dioxins: a 24-year follow-up study of the Yucheng cohort. *Diabetes Care*, 2008, vol. 31, no. 8, pp. 1574–1579. DOI: 10.2337/dc07-2449
84. Sergeev A.V., Carpenter D.O. Hospitalization rates for coronary heart disease in relation to residence near areas contaminated with persistent organic pollutants and other pollutants. *Environ. Health Perspect.*, 2005, vol. 113, no. 6, pp. 756–761. DOI: 10.1289/ehp.7595
85. Everett C.J., Frithsen I.L., Diaz V.A., Koopman R.J., Simpson W.M. Jr., Mainous A.G. 3rd. Association of a polychlorinated dibenzo-p-dioxin, a polychlorinated biphenyl, and DDT with diabetes in the 1999–2002 National Health and Nutrition Examination Survey. *Environ. Res.*, 2007, vol. 103, no. 3, pp. 413–418. DOI: 10.1016/j.envres.2006.11.002
86. Ha M.H., Lee D.H., Son H.K., Park S.K., Jacobs D.R. Jr. Association between serum concentrations of persistent organic pollutants and prevalence of newly diagnosed hypertension: results from the National Health and Nutrition Examination Survey 1999–2002. *J. Hum. Hypertens.*, 2009, vol. 23, no. 4, pp. 274–286. DOI: 10.1038/jhh.2008.124
87. Uemura H., Arisawa K., Hiyoshi M., Kitayama A., Takami H., Sawachika F., Dakeshita S., Nii K. [et al.]. Prevalence of metabolic syndrome associated with body burden levels of dioxin and related compounds among Japan's general population. *Environ. Health Perspect.*, 2009, vol. 117, no. 4, pp. 568–573. DOI: 10.1289/ehp.0800012

88. Lee D.H., Lee I.K., Porta M., Steffes M., Jacobs D.R. Jr. Relationship between serum concentrations of persistent organic pollutants and the prevalence of metabolic syndrome among non-diabetic adults: results from the National Health and Nutrition Examination Survey 1999–2002. *Diabetologia*, 2007, vol. 50, no. 9, pp. 1841–1851. DOI: 10.1007/s00125-007-0755-4
89. Park S.K., Son H.K., Lee S.K., Kang J.H., Chang Y.S., Jacobs D.R., Lee D.H. Relationship between serum concentrations of organochlorine pesticides and metabolic syndrome among non-diabetic adults. *J. Prev. Med. Public Health*, 2010, vol. 43, no. 1, pp. 1–8. DOI: 10.3961/jpmph.2010.43.1.1
90. Lee D.H., Lee I.K., Jin S.H., Steffes M., Jacobs D.R. Jr. Association between serum concentrations of persistent organic pollutants and insulin resistance among non-diabetic adults: results from the National Health and Nutrition Examination Survey 1999–2002. *Diabetes Care*, 2007, vol. 30, no. 3, pp. 622–628. DOI: 10.2337/dc06-2190
91. Ronn M., Lind L., van Bavel B., Salihovic S., Michaelsson K., Lind P.M. Circulating levels of persistent organic pollutants associate in divergent ways to fat mass measured by DXA in humans. *Chemosphere*, 2011, vol. 85, no. 3, pp. 335–343. DOI: 10.1016/j.chemosphere.2011.06.095
92. Wolff M.S., Anderson H.A., Britton J.A., Rothman N. Pharmacokinetic variability and modern epidemiology – the example of dichlorodiphenyltrichloroethane, body mass index, and birth cohort. *Cancer Epidemiol. Biomarkers Prev.*, 2007, vol. 16, no. 10, pp. 1925–1930. DOI: 10.1158/1055-9965.EPI-07-0394
93. Elobeid M.A., Padilla M.A., Brock D.W., Ruden D.M., Allison D.B. Endocrine disruptors and obesity: an examination of selected persistent organic pollutants in the NHANES 1999–2002 data. *Int. J. Environ. Res. Public Health*, 2010, vol. 7, no. 7, pp. 2988–3005. DOI: 10.3390/ijerph7072988
94. Lee D.H., Steffes M.W., Sjodin A., Jones R.S., Needham L.L., Jacobs D.R. Jr. Low dose of some persistent organic pollutants predicts type 2 diabetes: a nested case-control study. *Environ. Health Perspect.*, 2010, vol. 118, no. 9, pp. 1235–1242. DOI: 10.1289/ehp.0901480
95. Vasiliu O., Cameron L., Gardiner J., Deguire P., Karmaus W. Polybrominated biphenyls, polychlorinated biphenyls, body weight, and incidence of adult-onset diabetes mellitus. *Epidemiology*, 2006, vol. 17, no. 4, pp. 352–359. DOI: 10.1097/01.ede.0000220553.84350.c5
96. Lind P.M., Lind L. Circulating levels of bisphenol A and phthalates are related to carotid atherosclerosis in the elderly. *Atherosclerosis*, 2011, vol. 218, no. 1, pp. 207–213. DOI: 10.1016/j.atherosclerosis.2011.05.001
97. Vena J., Boffetta P., Becher H., Benn T., Bueno-de-Mesquita H.B., Coggon D., Colin D., Flesch-Janys D. [et al.]. Exposure to dioxin and nonneoplastic mortality in the expanded IARC international cohort study of phenoxyherbicide and chlorophenol production workers and sprayers. *Environ. Health Perspect.*, 1998, vol. 106, suppl. 2, pp. 645–653. DOI: 10.1289/ehp.98106645
98. Shcherbatykh I., Huang X., Lessner L., Carpenter D.O. Hazardous waste sites and stroke in New York State. *Environ. Health*, 2005, vol. 4, pp. 18. DOI: 10.1186/1476-069X-4-18
99. Melzer D., Rice N.E., Lewis C., Henley W.E., Galloway T.S. Association of urinary bisphenol a concentration with heart disease: evidence from NHANES 2003/06. *PLoS One*, 2010, vol. 5, no. 1, pp. e8673. DOI: 10.1371/journal.pone.0008673
100. Weichenthal S., Mallach G., Kulka R., Black A., Wheeler A., You H., St-Jean M., Kwiatkowski R., Sharp D. A randomized double-blind crossover study of indoor air filtration and acute changes in cardiorespiratory health in a First Nations community. *Indoor Air*, 2013, vol. 23, no. 3, pp. 175–184. DOI: 10.1111/ina.12019
101. Allen R.W., Carlsten C., Karlen B., Leckie S., van Eeden S., Vedal S., Wong I., Brauer M. An air filter intervention study of endothelial function among healthy adults in a woodsmoke-impacted community. *Am. J. Respir. Crit. Care Med.*, 2011, vol. 183, no. 9, pp. 1222–1230. DOI: 10.1164/rccm.201010-1572OC
102. Langrish J.P., Mills N.L., Chan J.K., Leseman D.L., Aitken R.J., Fokkens P.H., Cassee F.R., Li J. [et al.]. Beneficial cardiovascular effects of reducing exposure to particulate air pollution with a simple facemask. *Part. Fibre Toxicol.*, 2009, vol. 6, pp. 8. DOI: 10.1186/1743-8977-6-8
103. Pope C.A. 3rd., Ezzati M., Dockery D.W. Fine-particulate air pollution and life expectancy in the United States. *N. Engl. J. Med.*, 2009, vol. 360, no. 4, pp. 376–386. DOI: 10.1056/NEJMsa0805646
104. Morishita M., Thompson K.C., Brook R.D. Understanding air pollution and cardiovascular diseases: is it preventable? *Curr. Cardiovasc. Risk Rep.*, 2015, vol. 9, no. 6, pp. 30. DOI: 10.1007/s12170-015-0458-1
105. Huang W., Wang G., Lu S.-E., Kipen H., Wang Y., Hu M., Lin W., Rich D. [et al.]. Inflammatory and oxidative stress responses of healthy young adults to changes in air quality during the Beijing Olympics. *Am. J. Respir. Crit. Care Med.*, 2012, vol. 186, no. 11, pp. 1150–1159. DOI: 10.1164/rccm.201205-0850OC

Nosov A.E., Baydina A.S., Ustinova O.Yu. Aerogenic pollutants as risk factors causing development of cardio-metabolic pathology (review). *Health Risk Analysis*, 2021, no. 4, pp. 178–190. DOI: 10.21668/health.risk/2021.4.20.eng

Received: 27.09.2021

Accepted: 12.10.2021

Published: 30.12.2021